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Radon Concentration Due To Alpha Contribution Effects Of Soil And Rock Samples In Different West And Midlibyan Regions

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Abstract: The activity concentration of natural radioactivity for soil samples collected from western and mid Libyan regions were measured using HPGe detector. The average activity concentration of ^{226}Ra, ^{232}Th and ^{40}K for ten samples were found to be 51.86±7.14, 75.56±10.95 and 128.98±6.88 Bq/kg respectively. ^{235}U was detected for six samples with average activity 8.6±1.0 Bq/kg. The results obtained for the corresponding nuclides ^{226}Ra and ^{232}Th are above worldwide average values (35 and 30 Bq/kg) while ^{40}K was smaller than worldwide average (400 Bq/kg). The average outdoor absorbed dose and the annual effective dose rates due to ^{226}Ra, ^{232}Th and ^{40}K were observed to be 21.51±2.93 nGy/h and 0.297±.03 mSv/y respectively, which are lower than world average values (60nGy/h and 0.8 mSv/y). The radium equivalent activity and external hazard indices were found less than the world wide average values.

Keywords: Activity concentration, Absorbed Dose, Effective Dose, World average Activity values, Soil samples, Gamma spectroscopy, Western and Mid Libya

Introduction

The knowledge of radionuclides distribution and radiation levels in the environment is important for assessing the effects of radiation exposure due to both terrestrial and extraterrestrial sources. Natural background radiation is of terrestrial and extraterrestrial origin. Terrestrial radiation is due to radioactive nuclides present in varying amounts in rocks, building materials, water, soils and atmosphere. Natural radionuclides of uranium ^{238}U, thorium ^{232}Th and potassium ^{40}K are present in the earth's crust. When these radionuclides and their daughters in the series undergo decays gamma rays, beta and alpha radiations are released to the environment. Therefore, human beings are continuously exposed to ionizing radiation both inside and outside their dwellings. People ingest and inhale radionuclides through their food, air and water. The gamma ray exposure in room is due to radiation emitted decay products of ^{226}Ra, ^{232}Th series and ^{40}K. Human has always been exposed to natural radiation arising from the earth as well as from outside the earth [1,2 and 3] Natural environmental radioactivity and the related external exposure due to gamma radiation depend mainly on the

geological and geographical conditions, and appear at different levels in the soils of each region in the world. Every building construction material contains different quantities of natural radioactive nuclides. Radiation exposure due to building materials can be divided into external and internal exposure. The external exposure is caused by direct gamma radiation whereas internal exposure is caused by the inhalation of radon (^{222}Rn), thoron (^{220}Rn) and their short lived decay products. As, radon is a noble gas, it can transport easily through porous media for instance building materials, while usually only a fraction of that produced in the material reaches the surface and enters the indoor air .

The natural radioactivity in the environment is the main source of radiation exposure for human body. Natural radionuclide in soil contributes a significant amount of background radiation exposure to the population through inhalation and ingestion. It can be also transferred to plants and foods and drinking water.

According to A.L.A.R.A, principle, the radium equivalent Ra_{eq}, the external hazard index H_{ex}, the Absorbed Dose Rate A.D.R and the Annual Effective Dose Equivalent A.E.D.E were estimated and compared with results of other studies and with the worldwide average value.

This work was undertaken to measure the activity concentrations and γ-ray absorbed doses of the naturally occurring radionuclides in soil samples. Another aim of this work is to create the public awareness about the radiation hazards and it will also be helpful to establish a research base line in the investigated regions. Some previous studies of Libyan soil, Arab and neighbors' countries soil are summarized and tabulated in Table (1).

The samples were collected, as shown in map represented, in **Figure (1)**, at depth 1-10cm. Samples were treated thermally at 60 ^{0}C for 24 hours after that they were sieved to obtain uniform particle size about 550μm then the soil will filled in beaker which was sealed and then weighted and stored for a month to each the secular equilibrium.

Table. 1 Represents the mean and range of activity concentrations in Bq/kg of radionuclides for different countries.

Item	Mean activity concentration Bq/kg (range)			
	^{226}Ra	^{232}Th	^{40}K	country
1	7.5(4-13.5)	4.2(2.8-6.7)	27.5(19-39.6)	Libya, [4] Beach sand of Tripoli -2008
2	(58.8-102.1)			Libya, [5] Aljabel.Al.Gharbi-2011 Using Portable Nuclear Radiation Detector
3				Libya,[6]Northwestern,2001
4	26.02 (17.09-34.03)	30(11-64)	400(140-850)	Iraq, [7]kurkukoil field 2006
5	20.05 (10.87-30.94)	16.43(6.78-20.61)	216.69(127.74-272.7)	Qatar state [8]soil in Dukhan oil field 2015

6	22.03 (1.8-76.4)	27.91(6.3-85.5)	285.0(84-516.7)	Jordan, [9]soil in Tafilla2012
7	16 (46-115)	10	370	Kuwait[10]before war After war
8	(32.2-63.7) (5-13.8)	(44.3-95.6) (2.3-15.3)	(96-102) (29-582)	Egypt, [11],Beach sand Dune
9	30(2-110)	25(2-140)	370 (66-1150)	Algeria,[12]
10	23 (10-64)	20 (10-32)	270 (78-780)	Syria , [13],phosphate rocks
11	12.9 (0.15-41)	6.98 (4.21-9.92)	278 (167.79-419.52)	Nigeria, [14],Akuk,Ondo state

Figure 1: Distribution of selected Samples on Map of Libya

MATERIALS AND METHODS
Sample Collection and Preparation

The samples are selected from different geographic and geological regions in western and mid Libya. Two samples were collected, with masses varied 0.3-0.53 kg, for each region with separate distance 20 km. The samples information is illustrated in Table 2.

.Table. 2 Samples information

Samples I.D	Region	Weight (gm)	Type of sample	Sample Location	
				Latitude	Longitude
SA1	Qarabulli	330	Coast	32^0 $44^\backslash$	15^0 $14^\backslash$
SA2	Bu-njim	350	Sand (desert)	30^0 $35^\backslash$	15^0 $24^\backslash$
SA3	Zawia	480	Mountain(stones)	32^0 $45^\backslash$	12^0 $44^\backslash$
SA4	Qaddahea	510	Near the coast	31^0 $22^\backslash$	15^0 $14^\backslash$
SA5	Orban	450	Sand(Oises)	Near Ghrian	Near Ghrian
SA6	Tajoura	300	Sand (desert)	32^0 $53^\backslash$	13^0 $23^\backslash$
SA7	Sokna	530	Coast (near sea)	29^0 $10^\backslash$	16^0 $10^\backslash$
SA8	Ghrian	315	Coast	32^0 $21^\backslash$	15^0 $o8^\backslash$
SA9	Misurata	450	Mountain (stones)	32^0 $25^\backslash$	15^0 $05^\backslash$
SA10	Qaser Akhiar	240	Desert	Near Qarabulli	Near Qarabulli

Gamma-ray Spectrometry

Gamma spectrometry offers a convenient, direct, and non-destructive method to measure the activity of different radionuclides in the environmental samples. It also offers high efficiency Nal(Tl) detectors and high resolution (semiconductor detectors) detection. This technique enables the use of large quantities of samples to be counted. It is also possible, in this method, to reduce the essential background to very low values using suitable shielding arrangement. These advantages together with appropriate ability software (Genie 2000) that have now become available has made the gamma spectrometry method one of the most accurate technique for determining the activity concentration   of the environmental samples. In the present work, HPGe gamma spectrometer was used for the determination of gamma active radionuclide in soil and rock samples. NaI(Tl) scintillator and HPGe semiconductor detector are commonly used for the gamma ray spectrometry. Hyper pure germanium detectors are widely used for gamma ray spectroscopy to determine quantitatively the activities of natural ^{40}K, ^{232}Th, ^{226}Ra in the environmental samples. The HPGe detectors have very high resolution, but the efficiencies are low compared to those of scintillation detectors such as Nal(Tl). The high purity germanium detector can be produced from either n-type or p-type (Germanium) semiconductor material. The block diagram of HPGe gamma ray spectrometer system is shown in Figure (2).The spectrum was analyzed using multichannel analyzer (MCA) connected to computer using Genie-2000 software. The sealed sample was placed in the protection unit of gamma ray spectrometry for the counting time of six hours the energy resolution (FWHM) of this detector is 2.01 keV at 1.33 MeV (^{60}CO).
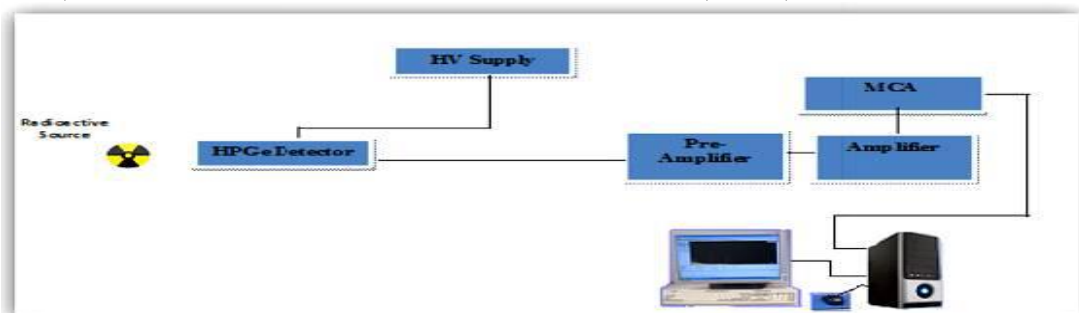

Figure (2): Block diagram of gamma-spectrometer detectors

Calibration of Gamma-ray Spectrometer System

The calibration of the spectrometer system for energy measurements is necessary to know the approximate energies of the radiation source being analyzed. The aim of calibration is to identify the radionuclide and activity concentrations present in an environmental sample. Energy calibration is carried out to confirm linear relationship between energy and the number of channels corresponding to that energy, and to determine the energy of each channel in a spectrum. The spectrum

is obtained for a reasonable time so that the photo peaks have sufficient counts for analysis. The regions of interest and centroid peak channel numbers are identified. Then the slope of the straight line plotted between the

Channel numbers versus energy represents the energy calibration factor. Figure (3) shows the energy calibration curve for HPGe detectors.
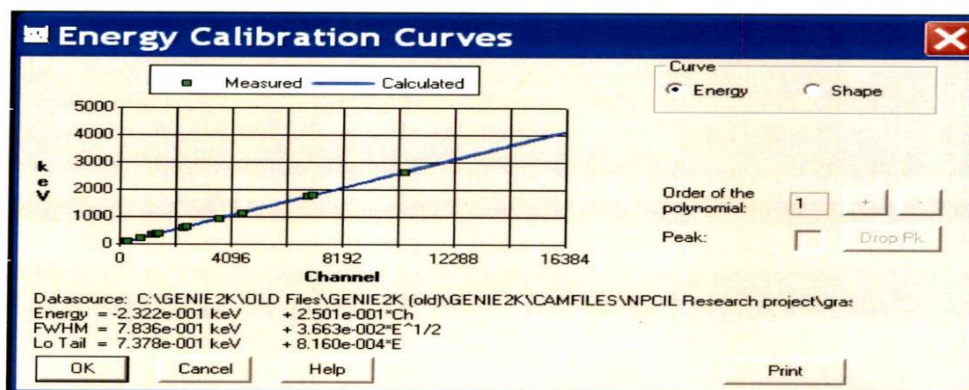

Figure (3): Energy calibration curve for the p-type HPGe detector generated by the Genie-2000 software.

In the present work, the detector efficiency calibration was performed using standard assurance reference materials and standard soil. These standard reference materials were taken in containers similar to the containers used for filling the soil samples for gamma spectrometric determination. The standard materials and samples were taken in containers of the same size and type so that the geometry remained the same. The samples were counted long enough (one week) to reduce the counting error. The variation of efficiency of the detector with energy for different gamma lines of various radionuclides for the p-type HPGe gamma ray spectrometer system used in the present work is shown in Figure (4).
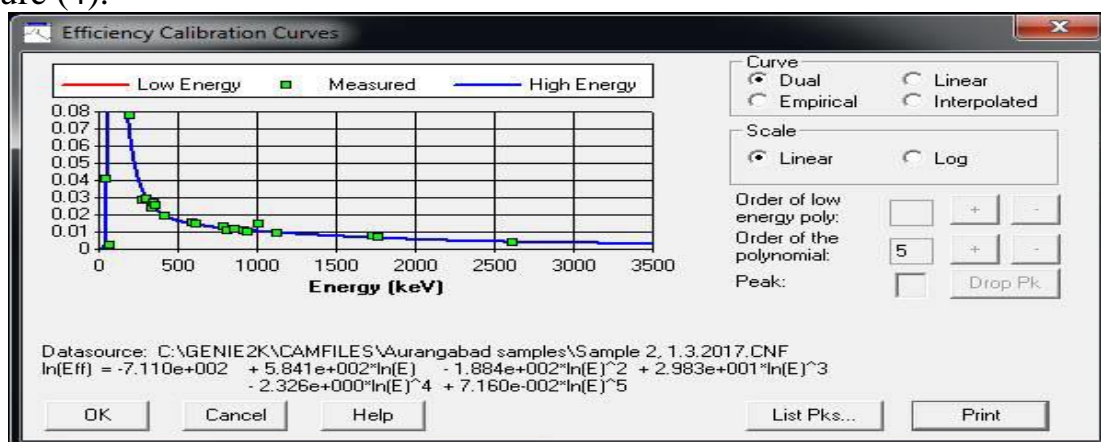
Figure (4): Efficiency calibration curve for the p-type HPGe gamma spectrometer system (as generated using Genie-2000 gamma spectrum analyses software).

Calculation of Activity Concentrations

The radioactivity of each sample was measured with keeping the samples one by one on the top of the detector and counted for a period of one day. The activity concentration (A) of each radionuclide in the sample was determined by using the count rates (N_c) (after subtracting the back ground).

$$A = \frac{N_C}{\varepsilon I_\gamma W} \qquad \textbf{(1)}$$

Where ε =Efficiency of the detector for the specific energy, I_γ= Intensity of the gamma ray and W= Sample weight (kg). For the analysis of peak areas of gamma spectra, a computer software programming (Genie 2000) was used [15]. Determination of NORM were carried out by measuring different daughters that emit clear gamma peaks of high intensity to confirm the attainment of radioactive secular equilibrium within the samples between ^{226}Ra and its daughters. This was carried out by measuring ^{226}Ra directly through the 186.2 keV and indirectly by measuring the ^{214}Bi (609.3, 1120.2 and 1764.5 keV) and ^{214}Pb (351.9 keV) photo peaks. ^{235}U was determined directly through the 143.8 keV photo peak. ^{232}Th was determined through ^{228}Ac (911.2 keV), ^{212}Pb (238.6 keV after subtract 241.2 value) and ^{208}Tl (2614 keV) photo peaks, and estimation of ^{40}K through the 1460.8 keV photo peak. The main radionuclides and specifications are listed in Table (3).

Nuclide to be determined		Nuclide measured	Energy in keV	Probability in %	Half-life time
^{226}Ra		^{226}Ra	186.1	3.51	1600y
		^{214}Pb ^{214}Pb	295.22 351.1	18.15 35.1	26.8 mit
		^{214}Bi ^{214}Bi	609.31 1120.22	44.6 14.7	19.9 min
^{232}Th	^{228}Th	^{228}Ac ^{228}Ac ^{228}Ac	209.25 338.32 911.2	3.89 11.27 25.8	6.13 h
	^{228}Ra	^{208}Tl ^{208}Tl ^{208}Tl	583.19 860.56 2614.53	30.4 4.47 35.64	307 min
^{40}K		^{40}K	1460.83	10.67	1.28×10^9 y

Table(3) : Natural Radionuclides found in samples and background [16].

Computation of Radiological Effects
Radium equivalent Activity Ra$_{eq}$

The important radionuclides in nature ^{226}Ra, ^{232}Th and ^{40}K are not uniformly distributed ,this due to disequilibrium between ^{226}Ra and its decay products .For

uniformity in exposure, estimates the radionuclide concentrations have been defined in terms of radium equivalent activity (Ra_{eq} in Bq/kg) . This allows comparison of the specific activity of materials (A_{Ra}, A_{Th} and A_K) containing different amounts of ^{226}Ra, ^{232}Th and ^{40}K according to:

$$Ra_{eq}=A_{Ra} + 1.43A_{Th} +0.077A_K \quad (2)$$

External and Internal Hazard Index:

The hazard index (H_{ex}, H_{in}) is the indoor radiation dose rate due to the external/internal exposure gamma radiation construction materials which was calculated by:

$$H_{ex}= 0.0027A_{Ra} +0.0038 A_{Th} + 2.o8\times10^{-4}A_K \quad \textbf{(3)}$$
$$H_{in}=0.00541 A_{Ra} +0.0038 A_{Th} + 2.o8\times10^{-4}A_K$$

Calculation of air absorbed dose rate:

The external outdoor absorbed gamma dose rates due to terrestrial γ-rays from the nuclides of ^{226}Ra, ^{232}Th and ^{40}K at 1m above the ground level was calculated as :-

$$A.D.R=(0.461A_{Ra} +0.623A_{Th} +0.0414A_K) \, nGyh^{-1} \quad (4)$$

About 98% of the external γ dose rate from ^{238}U series is delivered by the ^{226}Ra sub series .So disequilibrium between ^{226}Ra and ^{238}U will not affect the results of dose calculations from the measurement of ^{226}Ra**.** The absorbed dose rate was converted into annual effective dose equivalent by using conversion factor of o.7SvGy and 0.2 for the outdoor occupancy factor by considering that the people on the average spent 20% of the time outdoors.

Effective dose rates:

The Effective dose due to natural activity in soil was calculated by:
$$(A.E.D.E) \, in = 8760\times0.2\times0.7\times10^{-3}A.D.R) \, \mu Svy^{-1} \quad (5)$$

Gamma index (Iγ):

The index (Iγr) is used to estimate the level of γ –radiation hazard associated with the natural radionuclides in specific investigated samples, is defined as :-

$$I_{\gamma r} =0.007A_{Ra}+0.01A_{Th} +6.6\times10^4A_K \quad (6)$$

For materials that are used in bulk quantities (Such as clay bricks and concrete etc.), the value of Iγ ≤ 0.5 corresponds to a dose rate criterion of 0.3 mSv yr-1 whereas 0.5 < Iγ ≤ 1 corresponds to a criterion of 1 mSv y^{-1} [17]

Alpha index (Iα):

As radon daughters decay, they emit radioactive α- particles and attach to aerosols, dust and other particles in the air. As persons inhale, radon progeny are deposited on the cells lining the airways where the α- particles can damage DNA and potentially cause lung cancer. The excess α- particles radiation due to radon inhalation originating from building materials is estimated through the α- index (**Iα**), which is defined as follows [14]: Iα = $\dfrac{A_{Ra}}{200}$ (7) The recommended upper limit concentration of ^{226}Ra is 200 Bq kg^{-1} which gives Iα= 1.

Results and discussion

In this work the activity concentrations and radiological indices of soil samples, taken from different Libyan sites, are summarized. Activity concentrations for nuclides ^{235}U, ^{226}Ra, ^{232}Th and ^{40}K was determined by equation (1) and the results were tabulated in Table (4) and illustrated togather in Figure (5). The highest value is found for sample S10A, for ^{238}U (^{226}Ra) 103.81 Bq/kg, while SA4 for ^{232}Th(153.5Bq/kg) also SA4 for ^{40}K 168.57 Bq/kg). The highest activity, of the nuclide, may vary from place to place due to chemical changes in elements of soil or using agriculture fertilizers or material wastes, weapons etc . The results for these nuclides are also shown independently through Figures. (6-8).

Table (4) The Activities concentrations of the ^{235}U , ^{226}Ra , ^{232}Th and ^{40}K in Bq/kg for the measured samples.

| I.D	^{235}U	^{226}Ra	^{235}U/^{238}U	^{232}Th	^{40}K	^{226}Ra/^{232}Th
S1A	12.33±1.3	74±7.85	0.17	63.00±6.70	146.2±15.16	1.17
S2A	7.47.±0.82	56.83±6.77	0.14	42.96±4.44	131±13.39	0.78
S3A	5.32±0.66	33.06±4.53	0.16	40.74±4.94	146.4±16.2	2.13
S4A	10.79±1.65	60.16±6.52	0.18	153.5±16.47	168.57±17.30	1.56
S5A	7.83±0.76	29.12±5.42	0.27	92.33±8.79	108.57±11.14	1.09
S6A	7.89±0.78	30.09±3.86	0.26	102.3±9.02	116.01±13.24	0.57
S7A	N.D	26.71±5.95	N.D	47.62±6.21	103.44±10.36	1.42
S8A	N.D	33.26±4.49	N.D	46.92±6.36	100.35±9.65	0.71
S9A	N.D	71,6±7.08	N.D	47.7±6.19	140.3±14.52	2.25
S10A	N.D	103.81±11.28	N.D	118.3±11.60	N.D	0.87
Average	8.6±1	51.86±5.14	0.196	75.56±7.36	128.98±12.05	1.305
Max	12.33	103.81	0.27	153.5	168.57	2.25
Min	5.32	26.71	0.14	40.74	100.35	0.57
S.D	0.99	5.14	0.0005	7.36	12.05	0.16

N.D Not Detected
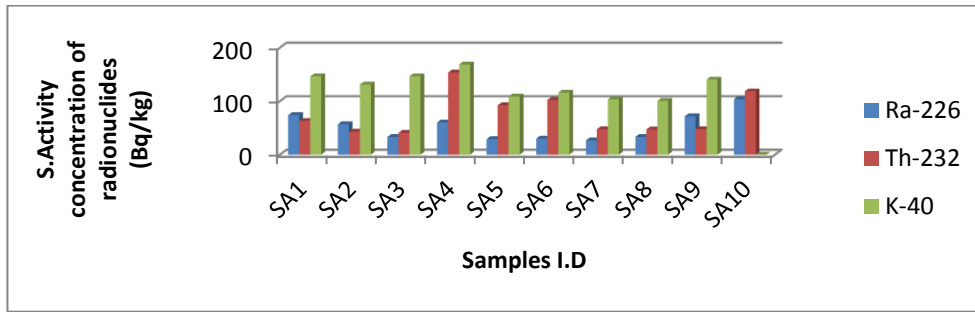
Figure (5):The activity concentration of radionuclides for investigated samples

From Table (3), ^{235}U appears in six regions (labeled from 1 to 6 as illustrated in Figure (6) with high abnormal concentrations (5.32-12.33 Bq/kg) and higher ratio (o.14-0.27) than normal (0.026) as illustrated in Figure (7), the samples were taken from these regions after NATO strikes during sixth months 2011; the type of used weapons has fission nature.
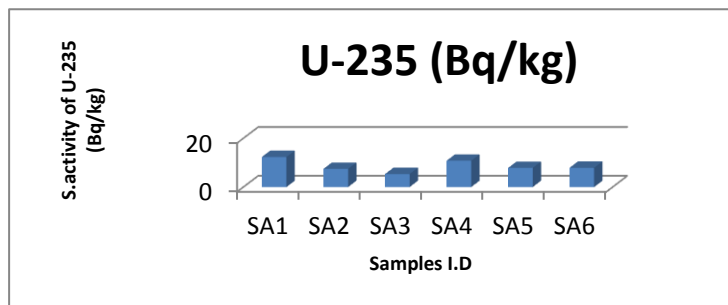

Figure (6): Soil Samples Activity Concentration of ^{235}U
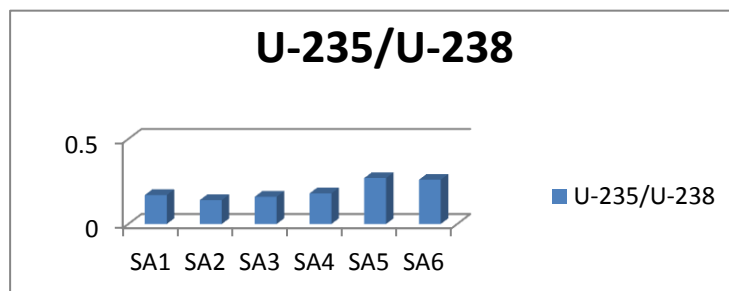
Figure (7) ^{235}U to ^{238}U activity concentration ratio $^{226}Ra\ activity$

In soil samples the activity concentrations of ^{226}Ra were found in the range of 26±0.71-103.8±12.38Bq/kg, with an average value 51.68±7.14 Bq/kg .This result is higher than the world wide average value of 35 Bq/kg for the same radionuclides in soils reported by UNSCEAR.

Ra-226 (Bq/kg)
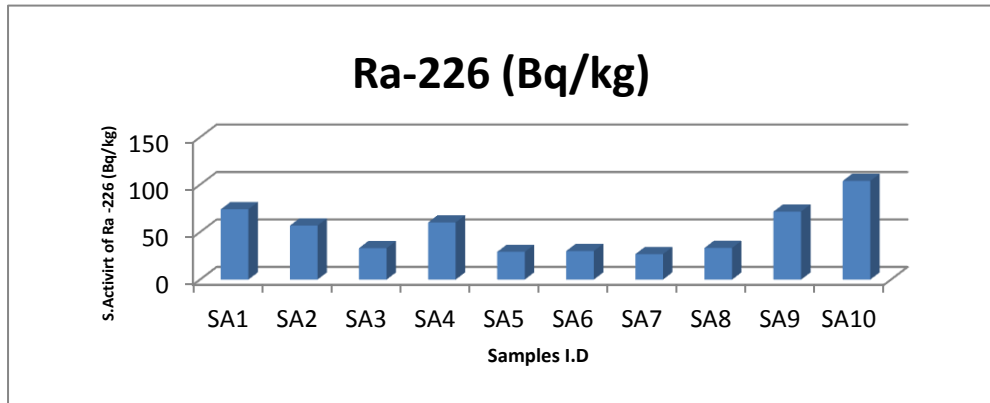
Figure(8): Soil Samples versus activity concentration of ^{226}Ra radionuclides (^{238}U Series) ^{232}Th Activity

The concentration of ^{232}Th is found in the range 40.74 ± 2.70 - 153.5 ± 7.57 Bq/kg with mean value 75.56 ± 10.95 Bq/kg this result is higher than the world average of 30 Bq/kg. SA1 Qarabulli).The ratio concentration of ^{226}Ra(^{238}U series) to ^{232}Th is less than unity,(concentration of ^{232}Th is higher than ^{226}Ra)

Th-232
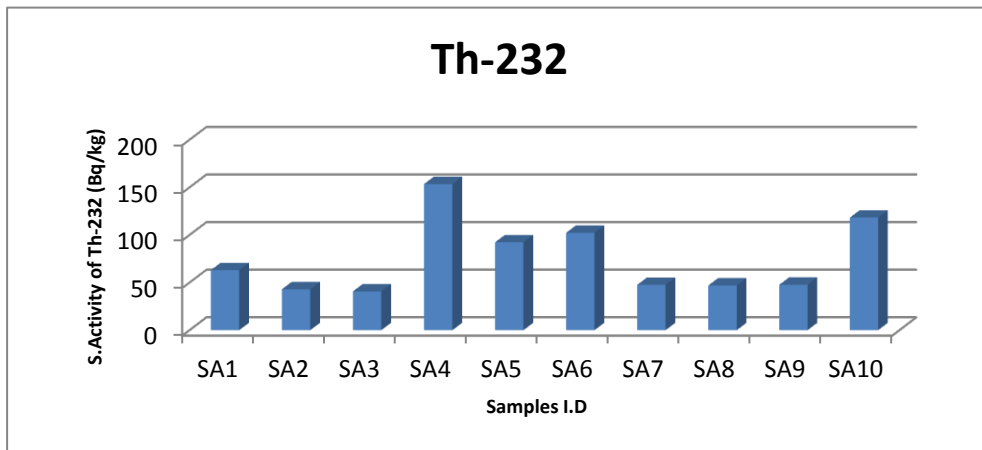
Figure (9): Soil Samples activity concentration of of ^{232}Th series ^{40}K Activity

The activity ^{40}K is found in the range (100.35 ± 4.36) – (168.57 ± 27.26) Bq/kg with the average value of 168.98 ± 6.88 Bq/kg . This result is mostly lower than the world wide average of 400 Bq/kg but SA4 is very high for the same kind of nuclide.

K-40
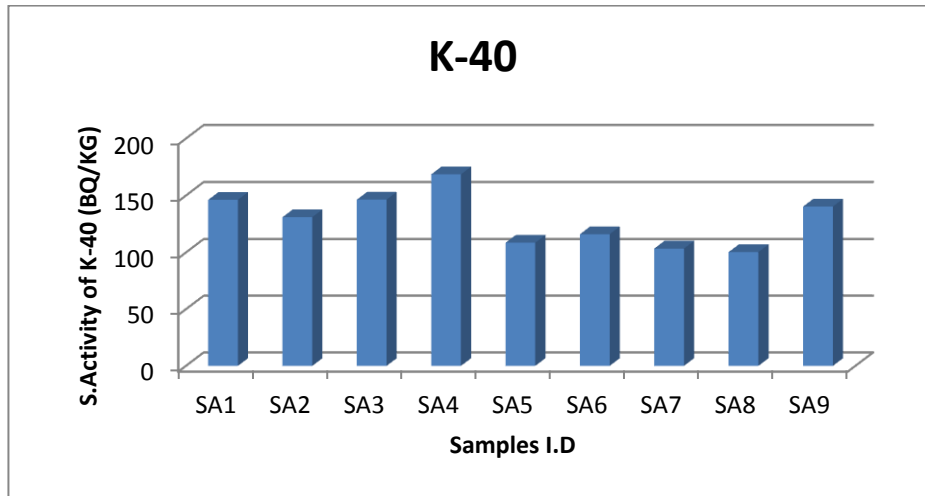
Figure (10): Soil Samples concentration of ^{40}K

Ra-226/Th-232
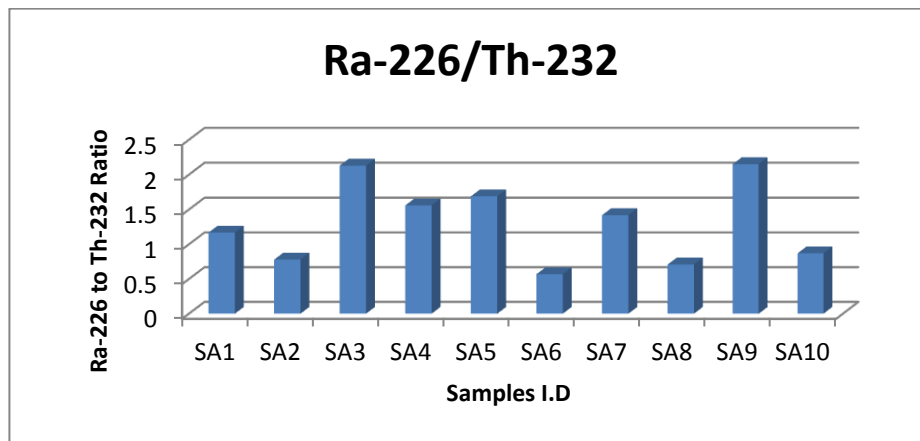
Figure (11): ^{238}U(^{226}Ra) to ^{232}Th Activity Ratio

To estimate the health effects ,the radiation hazards such as radium equivalent (Ra$_{eq}$),external and internal hazard index (H$_{ex}$,H$_{in}$) ,absorbed dose Rate (A.D.R) ,effective dose rate (AEDE), level index (I$_{\gamma r}$) and α- -index have been calculated from the activity of nuclides ^{226}Ra ,^{232}Th and ^{40}K using the equations (2-6) respectively and the values have shown in Table(5) and Figure.(12-16)
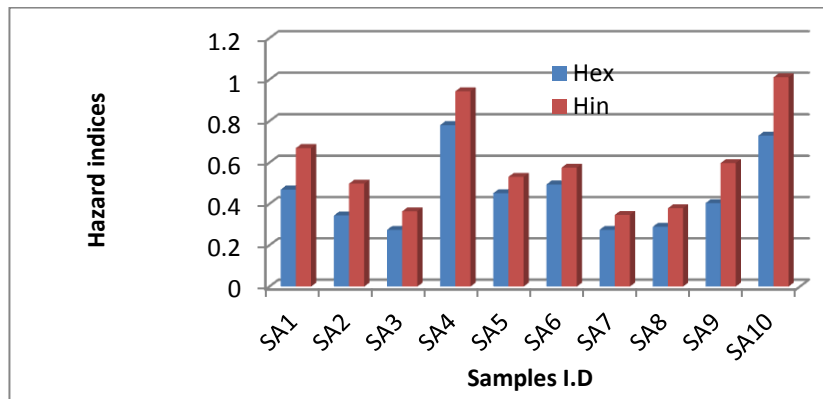
Figure.(12): Radiological Indices of the Investigated Sample (Hazards-Indices)
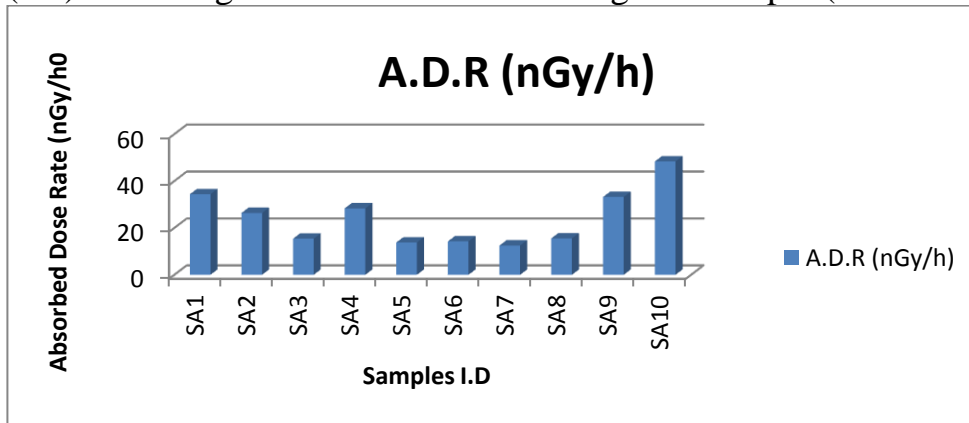
Figure.(13): Radiological Indices of the Investigated Sample (Absorbed Dose)
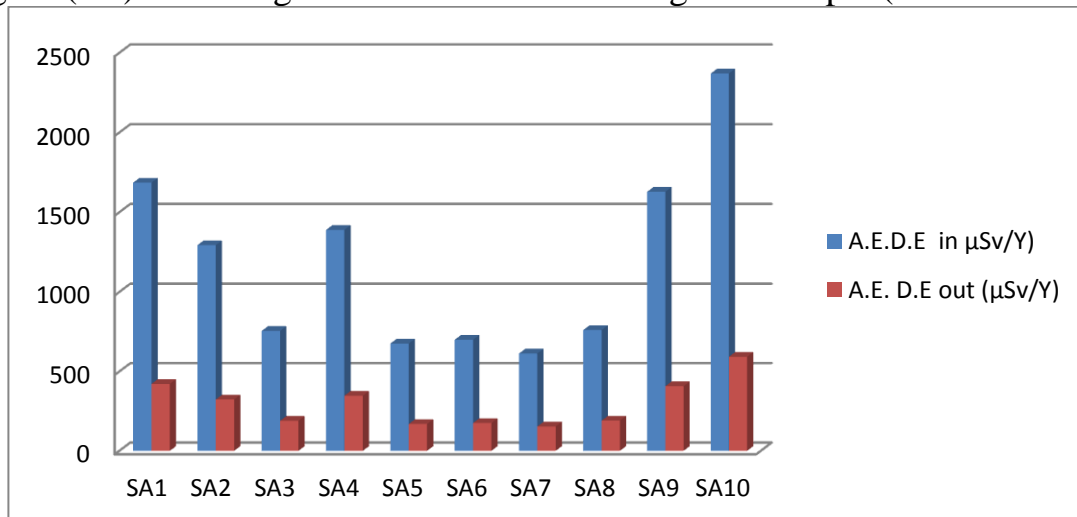
Figure.(14): Radiological Indices of the Investigated Sample (Annual Effective Dose)
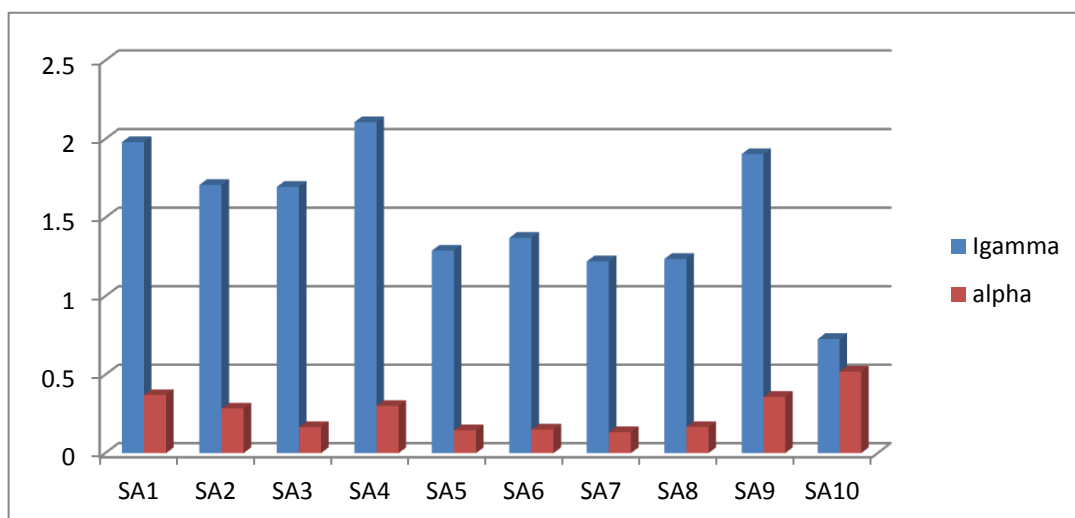
Figure.(15): Radiological Indices of the Investigated Sample ($I_{\gamma r}$ and I_α indices)

Table (**5**), shows that the radium equivalent (Ra_{eq}) is found in the range (102.59±20-292.64±34.33 Bq/kg), and has average value of 157.25±13.18 Bq/kg. The average value of radium equivalent is less than the safe limits 370 Bq/kg[2]. The mean value of external radiation hazard index is (0.45±0.09) which is less than 1 and confirm it as safe to carry out the activities for the human in that region . The outdoor air absorbed dose rate due terrestrial gamma rays at 1m above the ground were calculated for ^{226}Ra ,^{232}Th and ^{40}K and the range is (12.49±1.66) – (34.35±3.76) nGy/h with an average 21.52±2.20 nGy/h which is lower than the world average of 60 nGy/h[2].The annual effective dose rate equivalent is calculated using a conversion factor of 0.7 Sv/Gy to convert the absorbed dose rate to the effective dose equivalent and 0.2 for the outdoor occupancy factor. The annual effective dose rates are found in the range of (153.23±14.50 - 263.89±28.1) µSv/y with an average-e 296.75±29.60 µSv/y which is lower than the world average of 1000µSv/y for the general public (UNSCEAR, 2000).The representative level index $I_{\gamma r}$ equation (6) must be less than unity. For the investigated samples this index is in average 1.52±0.20 Bq/kg, where higher than unity in most samples. Table (5) gives Average activity concentration   compared   with   others.

Table(5). Radium Equivalent Activity (Ra_{eq}), Absorbed Dose Rate (A.D.R),External Hazard Index (H_{ex}), (A..E.D.E) and Representative level index ($I_{\gamma r}$).

I.D	Ra_{eq}(Bq/kg)	A.D.R (n Gy /y)	H_{in}	H_{ex}	A.E.D.E (µ.Ss/y)	$I_{\gamma r}$ (Bq/kg)	I_α Index
S.A1	175.34	34.35	0.67	0.469	421.31	1.980	0.37
S.A2	128.34	26.36	0.49	0.34	323.30	1.70	0.28

S.A3.	102.59	15.39	0.36	0.274	188.81	1.69	0.17
S.A4	292.64	28.31	0.94	0.78	347.28	2.10	0.30
S.A5	169.51	13.77	0.53	0.45	168.94	1.28	0.15
S.A6	185.31	14.26	0.57	0.49	174.89	1.37	0.16
S,A7	102.77	12.49	0.34	0.27	153.23	1.22	0.13
S.A8	108.08	15.51	0.37	0.28	190.23	1.23	0.18
S.A9	150.61	33.18	0.59	0.40	407.03	1.90	0.36
S.A10	272.97	48.30	1.011	0.72	592.42	0.72	0.52
Average	157.247	21.51	0.59	0.45	296.74	1.52	0.25
Max.-	292.64	34.35	1.01	0.78	263.89	2.10	0.52
Min.	102.591	12.49	0.34	0.27	153.23	0.72	0.13
S.D	16.05	2.93	0.023	0.027	0.030	0.43	0.096

Table (6) Average activity concentration in this work and others (Bq/kg)

Country	^{238}U	^{232}Th	^{40}K	Ref.
Malizia	39±0.7	52±1	61±15	[18]
Egypt	79±2	44±1	586±18	[19]
Libya	7.5±2.5	6.7±1.9	4.5±1.3	[20]
Iran	74±4	69±4	1130±32	[21]
Turky	70±0.8	83±1	1234±7	[22]
Kenya	12.63-72.51	11.45-58.12	234.8-1058.52	[23]
Nigeria	74.74±5.67	199.23±43.30	1021.27±7.14	[24]
This work	51.86±4.38	75.56±7.36	128.98±12.09	
World wide	33	45	412	[2]

Correlation study:

Correlation analyses were performed to reveal the possible relationship between concentrations of different radionuclides in the samples. The Pearson product-moment correlation matrixes for the correlation coefficient values (R) between the radionuclides activity concentrations were calculated. The correlation between ^{226}Ra &^{232}Th, ^{226}Ra &^{40}K and Radium and absorbed dose in air of soil samples is computed from the concentrations of these radionuclides and shown in figures (16 to 18) respectively. There is a weak correlation between^{226}Ra and ^{232}Th, for the samples (with negative correlation coefficients R= 0.504) in all sampling locations. And there is negative correlation between ^{226}Ra &^{40}K[Correlation coefficient R = -1.43] in samples. The value of correlation between ^{226}Ra and absorbed dose significantly higher with positive correction [correction coefficient R=0.99]
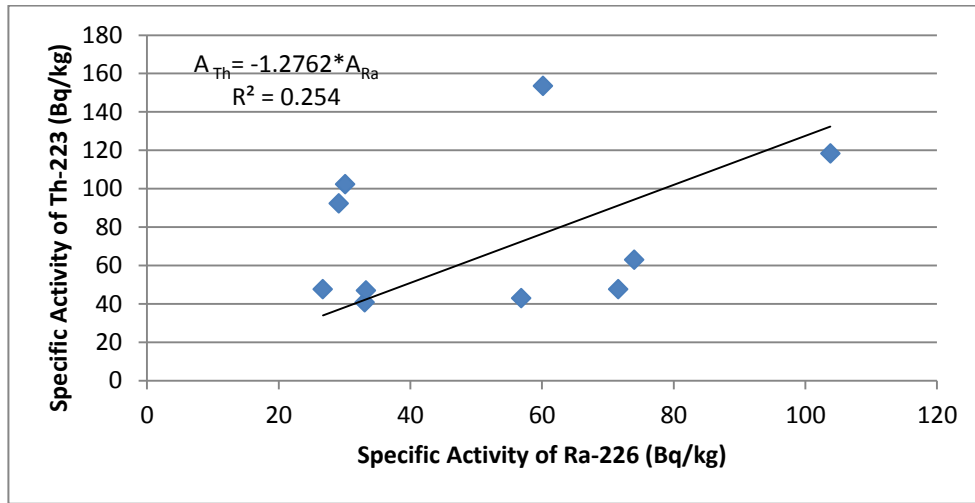
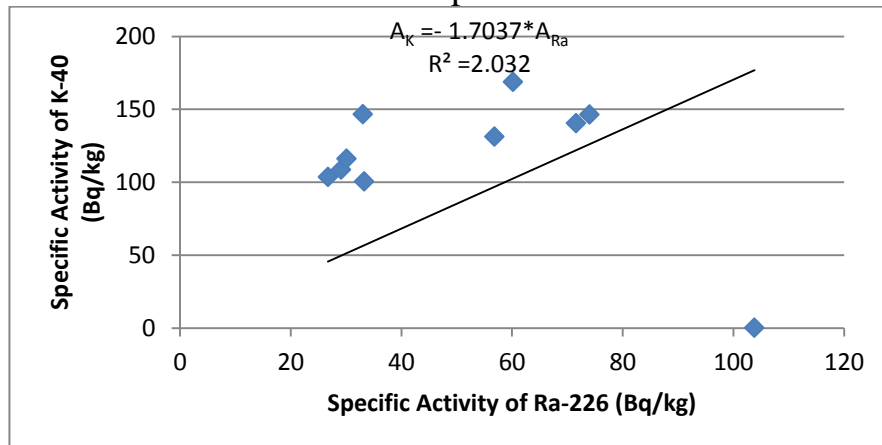Figure (16): Correlation between Radium and Thorium concentration in Soil samples

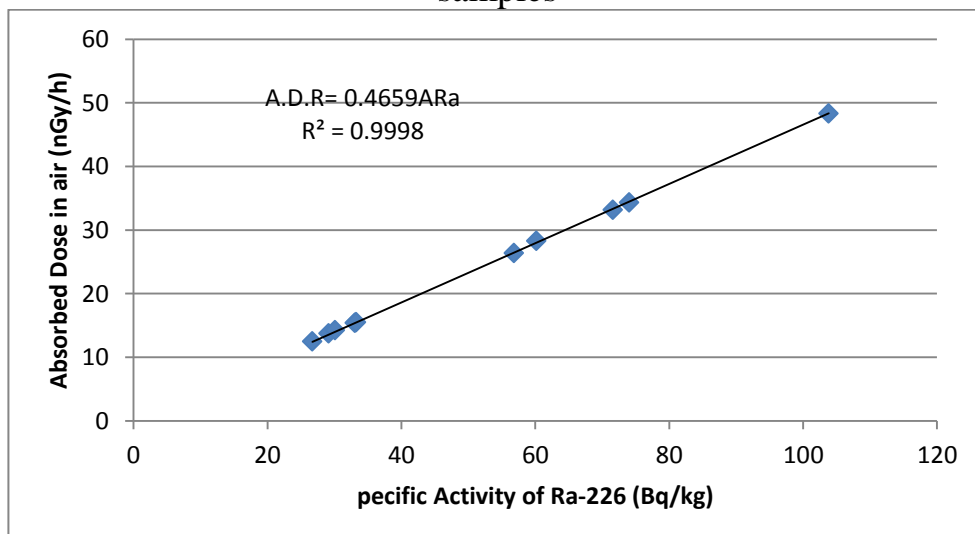Figure (17) Correlation between Radium and Potassium concentration in Soil samples

Figure (18): correlation between ^{226}Ra and absorbed dose in air

Conclusion

The activity concentration of ^{226}Ra, ^{232}Th and ^{40}K has been measured for some soil and rock samples from different locations of West and Mid Libya by using gamma-ray spectrometry (HPGe) detector. The activity of ^{226}Ra, ^{232}Th and ^{40}K in samples are found to ranges from 26.71± 5.95 to 103.31± 11.28 Bq kg^{-1}, 40.74±4.94 to 153.5±16.47 Bq kg-1 and100.35±9.65 to 168.57±17.30 Bq kg-1, respectively. While world average concentrations are 35, 30 and 400 Bq kg-1 for ^{226}Ra, ^{232}Th and ^{40}K, respectively (UNSCEAR, 2000). The average and ranges of activity concentration of ^{226}Ra , ^{232}Th in soil of these areas are quite higher than the world average reported values (UNSCEAR, 2000) while for ^{40}K less than world range . The Average value of radium equivalent activity is 157.25 Bq kg-1 which is below the recommended value of 370 Bq kg-1. T he values of absorbed dose rates due to ^{226}Ra, ^{232}Th and ^{40}K in soil samples vary from 12.47 to34.35 nGy h-1 with an average value of 21.52 nGy h-1 . The calculated values of absorbed dose has been found lower than global average value. The annual effective dose rates in outdoor is found to vary from 0.153 to 0.264 mSv y-1with an average value of 0.297 mSv y-1. This is below the limit of 1 mSv y-1 for general population (UNSCEAR, 2000). The calculated values of external hazard Hex are vary from 0.27 to 0.78 with an average value of 0.45 whereas internal hazard index Hin are vary from 0.34 to 1.01 with an average value of 0.59. All values of Hex and Hin are less than unity except SA10 for H_{in} . However, the value of gamma index Iγ is found to vary from 0.71 to 2.6 with an average value of 1.52 and most values of Iγ were also found higher than one. All the values of Alpha index Iα were found below the maximum permissible value i.e.1.

For samples SA1, SA6,SA8 and SA10 the ratio of activity concentration (^{226}Ra/^{232}Th) is less than unity; this is due to that monazite contains more thorium than uranium. Also the abnormal activity concentration and ratio activity concentration (^{235}U/^{238}U) that is more than normal (0.026), this phenomena can be explained due to NATO strikes war during 2011, at these regions. The obtained result in this work can be used as the regional base line data for estimation the future radioactivity contamination in the studied regions.

Therefore, the soil of some regions used in the present study is exempted from all the restrictions concerning radioactivity, also these soil samples are safe to be used in building construction. However the soil samples of regions as mentioned above have higher values for ^{232}Th, ^{226}Ra, ^{235}U and radionuclides indices than world (UNSCEAR, 2000)[25]. The mean value of gamma index is obtained above the limit of 1 for most samples. On the basis of these results, researcher concluded that the soil of the study area (in particular regions) have radiological health hazard to the public.

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الفهـــــــرس

324–331	Hosam Ali Ashokri Fuad Faraj Alamari	The Inhibitory Effect of Common Thyme Thymus vulgaris Aqueous Extracts on Some Types of Gram-Positive and Gram-Negative Bacteria that Infect the Human Respiratory System	16
332-348	إنتصار علي ارهيمة	استخدام تحليل التباين الأحادي (لدراسة تأثير الملوحة على نبات الشعير)	17
349-363	إنتصار احمد احميد	ميناء الخمس البحري	18
364-386	فرج محمد صالح الدريع	تجار ولاية طرابلس الغرب والتغير في السلع (دراسة وثائقية في أحد مصادر تكوين الثروة) (1835–1912م)	19
387-413	حنان علي محمد خليفة	" قضية الإلهام في الشعر "	20
414-427	أحمد على معتوق الزائدي	الرجل المحرم للمرأة في الشريعة الاسلامية	21
428-447	محمد عبد السلام دخيل عبد اللطيف سعد نافع	الثقافة الاستهلاكية عند الشباب في ليبيا "دراسة ميدانية في مدينة الخمس"	22
448-471	إلهام نوري الشريف نورية محمد أبوشرنته	النظام الانتخابي في ليبيا عام 2012م	23
472-487	Salem Mohamed Edrah Afifa Milad Omeman	The Phytoconstituents Screening and Antibacterial Activities of Leaves, Seeds Bark and Essential Oil Extracted from Carya illinoinensis Plant	24
489-505	أحمد المهدي المنصوري	النص الشعري بين التأويل والتلقي خطاب الصورة عند الرقيعى أنموذجاً	25
506-521	Ibrahim M. Haram Mohamed E. Said Ahmad M. Dabah Osamah A. Algahwaji	Energy Recovery of Ethylene Dichloride (EDC) Production by Pinch Analysis (Abu-Kamash EDC plant)	26
522-544	زهرة المهدي أبوراس هنية عبد السلام البالوص	التنمر المدرسي بين الطلاب تعريفه ، أسبابه، أنواعه ومخاطره، وطرق مواجهته وعلاجه	27
545-565	عبدالله محمد الجعكي	حذف المفعول به اقتصارا واختصارا دراسة نحوية دلالية تطبيقية في نماذج من شعر ابن سنان الخفاجي	28
567-579	Najah Mohammed Genaw Sahar Ali Aljamal	EFL Learners' Attitudes towards the Use of Vocabulary Learning Strategies	29
580-592	نور الدين سالم رحومة قريب مسعودة رمضان علي العجل	الزمان الوجودي عند هيدجر وعبد الرحمن بدوي	30
593-600	Rajaa Mohamed Sager Saeeda Omran Furgan	Study of the relationship between the nature of wells water in Libyan southwestern zone and the occurrence of corrosion in the transferring metal pipelines	31

601-616	Sami Muftah Almerbed Abdumajid Mohamed Haddad Milad Ali Abdoalsmee	Evaluation of the Use of Technology in Private Schools	32
617-630	اسامة عبد الواحد البكوري ريم فرج بوغرارة	(جماليات الضوء في فن النحت) (دراسة تحليلية)	33
631-640	Affra A B Hemouda Silla Hiba Abdullah Ateeya Abdullah	Modern Technology in Database Programming, Software Engineering in Computers	34
641-656	Ashraf M. Saeid Benzrieg Abdullah M. Hammouche Abdelbaset M. Sultan	Prediction of Chronic Kidney Diseases Using Artificial Neural Network	35
657-674	Abdu Assalam A. Algattawi Ali M Elmansuri	Radon Concentration Due To Alpha Contribution Effects Of Soil And Rock Samples In Different West And Midlibyan Regions	36
675-692	Mohamed Ali Abunnour Nuri Salem Alnaass Mabruka Abubaira	Demographic Analysis of Socioeconomic Status and Agricultural Activities in Sugh El-Chmis Alkhums 1973-2014	37
693-704	Abdulbasit Alzubayr Abdulrahman Omar Ismael Elhasadi Zaynab Ahmed Khalleefah	Some applications of harmonic functions	38
705-729	عبدالحميد مفتاح أبو النور حنان فرج أبو علي محمد ابوعجيلة البركي	استشراف المستقبل و توظيف التطبيقات الالكترونية الذكية في تعليم تلاميذ مرحلة التعليم الاساسي	39
730-756	رجعة سعيد محمد الجنقاوي عبدالسلام ميلاد المركز	الاستهلاك المائي في منطقة الخمس ومشكلاتها والبدائل المطروحة لحلها	40
757-773	سيف بن سليمان بن سيف المنجي سماح حاتم المكي محمد رازمي بن حسين	التعلم عن بعد في حالات الطوارئ: تطبيقات التدريس وتجربة التعليم بمدارس التعليم ما بعد الأساسي في سلطنة عُمان	41
774-780	Aisha ALfituri Benjuma Najmah ALhamrouni Ahmed	Estimation of lead (II) concentration in soil contaminated with sewage water of Alkhums city	42
781-786	Hanan Saleh Abosdil Rabia Omar Eshkourfu Atega Said Aljenkawi Aisha Alfituri Benjuma	Determination of Calcium in Calcium Supplements by EDTA Titration	43
787-805	ميسون خيري عقيلة ابوبكر محمد محمد عيسي	مستوي القلق وعلاقته بالغربة عن الذات	44